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Impact of the soft-shell clam *Mya arenaria* on sulfate reduction in an intertidal sediment

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ABSTRACT: Sulfate reduction and various parameters related to the sulfur cycle were examined at mm to cm scales around burrows of the soft-shell clam *Mya arenaria* in an intertidal sediment (Lowes Cove, Maine, USA). Sulfate reduction rates were 1.5 to 2 times higher in the inner 1 to 5 mm region surrounding the burrow than in ambient sediment. In contrast, pools of reduced sulfur increased with the distance from the burrow wall to values ~1.5 times higher in ambient sediment. The highest numbers of sulfate-reducing bacteria (estimated using a most-probable-number technique) and microbial biomass (estimated from phospholipid phosphorous content) relative to ambient sediment were found in the innermost zone around burrows. Results from an artificial burrow experiment showed that artificial burrow irrigation suppressed sulfate reduction in the innermost zone around burrows, while radial profiles of reduced sulfur resembled those from *M. arenaria* burrows, indicating loss of reduced sulfur from the burrow wall. *M. arenaria* burrows are thus sites of enhanced microbial activity and a dynamic sulfur cycle, with turnover times of reduced sulfur compounds increasing with distance from the burrow wall. Enhanced sulfate reduction rates near burrows are likely caused by substrate enrichment, perhaps due to organic excretions from *M. arenaria*. The pattern of reduced sulfur turnover likely results from periodic oxygen inputs during burrow irrigation.

KEY WORDS: Bioturbation · Marine sediment · Sulfate reduction · Sulfate-reducing bacteria · *Mya arenaria*

INTRODUCTION

Benthic macrofauna profoundly affect the geochemistry and microbial processes of coastal sediments through feeding, burrowing and burrow ventilation (e.g. Aller 1982, 1988, Andersen & Kristensen 1988, 1991, Kristensen 1988). Sediment reworking buries fresh reactive organic matter from surface sediments to cm depths and exposes new surfaces for microbial colonization (e.g. Rhoads 1974, Yingst & Rhoads 1980). Burrow ventilation affects the exchange of solutes between sediment and overlying water, and is considered an important control of both aerobic and anaerobic microbial processes (Aller 1988, Kristensen 1988). The net result of macrofaunal activity is stimulated microbial growth and metabolism in the sedimentary environment. Accordingly, the burrows of infaunal macroinvertebrates are organic-rich sites of increased heterotrophic and chemolithotrophic activity (e.g. Aller & Yingst 1978, Kristensen et al. 1985, Aller & Aller 1986, Reichardt 1988).

Sulfate reduction generally predominates terminal carbon oxidation in coastal marine sediments, typically accounting for about 50% of total mineralization, and about 90% of the anaerobic mineralization of organic matter (e.g. Sørensen et al. 1979, Jørgensen 1982, Jørgensen & Sørensen 1985, Skyring 1987). However, despite the importance of sulfate reduction, relatively few studies have examined the role of benthic macrofauna on microbial processes related to the sulfur cycle. In a classic study, Aller & Yingst (1978) demonstrated elevated rates of sulfate reduction in the burrow wall of the polychaete *Amphitrite ornate*, and proposed that increased mineralization rates due to burrowing infauna would mainly be caused by a stim-
ulation of sulfate reduction in the sediment. Hines & Jones (1985) showed that macrofaunal activity caused higher sulfate reduction rates relative to a non-bioturbated sediment, and argued that sulfate reduction was enhanced due to continual mixing of the sediment and redistribution of fresh reactive organic matter from surface sediment to depth in sediment. Similar effects on sulfate reduction have been reported by Goldhaber et al. (1977) and Berner & Westrich (1985).

The soft-shell clam *Mya arenaria* L. occurs widely in the intertidal and subtidal zones of sub-boreal and boreal regions. This sessile, suspension-feeding bivalve is situated at 15 to 25 cm depth in the sediment, and accesses the surface by means of its fused inhalent and exhalent siphons. Previous studies have shown that populations of *M. arenaria* have important roles in benthic community metabolism and energy flow (Emerson et al. 1988). Filter-feeding activities of the clams also have a substantial role in transferring particulate material from the water column to the benthos (Nichols 1985). The impact of *M. arenaria* burrows on benthic nitrification has been investigated by Henrikson et al. (1983), who demonstrated increased potential nitrification rates in the oxidized zone surrounding the burrows compared to ambient sediment. More recently Pelegri & Blackburn (1995) showed that both nitrification and denitrification are enhanced by *M. arenaria*.

The objective of the present study was to investigate the impact of *Mya arenaria* and its burrows on sulfate reduction and other aspects of the sulfur cycle in an intertidal sediment. Sulfate reduction was measured directly along radial profiles around *M. arenaria* burrows at mm intervals. The observations from *M. arenaria* burrows were compared with results from an experimental laboratory study using artificial burrows.

**MATERIALS AND METHODS**

**Site description.** Sediment used in this study was obtained from Lowes Cove, an intertidal mudflat on the Damariscotta River estuary in Maine, USA (43° 56' N, 69° 35' W). The cove is approximately 600 m long and 150 m wide and has an area of about 5.9 ha. The mean tidal range is approximately 3 m and the salinity is constant around 28 to 30%. Water temperature varies between a maximum of about 20°C in mid-August and a low around 0°C in mid-February. The sediments of Lowes Cove, which consist of fine silts and clays with an organic content of 2 to 5% dry wt, is highly bioturbated by macrobenthos. The most common macrofauna are the bivalve molluscs *Macoma balthica*, *Mya arenaria* and *Mytilus edulis*, littorinid gastropods *Hydrobia* spp., and a wide variety of polychaetes (e.g. *Nereis succinea*, *Heteromastus filiformis*, *Scoloplos* sp.). Various aspects of sediment chemistry and biology at this site have been described previously (e.g. Anderson et al. 1981, King et al. 1983, Rice 1986). A sampling site was chosen at the head of the cove where high densities of *M. arenaria* occur (up to 200 ind. m⁻²; Rice 1986). On an average high tide, water depth at this site is approximately 1 to 2 m; the sediment is exposed for about 2 to 3 h during low tide.

**Sampling.** Sediments were collected at the onset of low tide when the mudflat was exposed. Sediment samples were obtained horizontally from *Mya arenaria* burrows after exposing a vertical profile of the sediment by digging with a 'clam fork'. During digging, the sediment typically cracked along burrow axes, exposing the burrow wall. A small undisturbed sediment sample was obtained by gently inserting a 5 cm³ 'cutoff' syringe 4 to 5 cm horizontally into the sediment (Fig. 1). Samples were obtained from burrows with a diameter of approximately 10 to 12 mm, and were taken 5 to 6 cm below the sediment surface. To avoid interference from neighbouring burrows, only burrows with no adjacent openings within a distance of 8 to 10 cm were sampled. The syringes were sealed with butyl septa immediately after sampling and were returned to the laboratory for further processing (within 30 min after sampling). *M. arenaria* burrows were sampled on 2 occasions, in June and November.
1994, respectively. Three additional sediment cores were collected in July 1994 with acrylic core liners (i.d. 2.9 cm) for the determination of vertical profiles of sulfate reduction in the upper 10 cm sediment. Samples were obtained at the same location as the 'Mya samples', but care was taken to avoid coring close to *M. arenaria* burrows. In situ water temperature ranged between 14 and 18°C. In the summer period, however, temperatures in the upper exposed sediment layer likely increased significantly during low tide due to high solar illumination.

**Sulfate reduction.** Radial profiles of sulfate reduction were made from syringe-collected sediment at mm intervals from the burrow wall and 25 mm into the ambient sediment, using a modified version of the \( {^{35}}S\)-SO\(_4^{2-}\) injection technique (Jørgensen 1978a). Sulfate reduction was determined in 9 burrow samples. To each syringe was injected 10 µl of carrier-free \( {^{35}}S\)-SO\(_4^{2-}\) (600 kBq) through the septum along the central axis using a Hamilton syringe that was slowly withdrawn through the sediment sample as the radiotracer was injected, thus ensuring as a distribution of \( {^{35}}S\)-SO\(_4^{2-}\) as possible. Samples were incubated in the dark for 4 to 5 h at 18°C. Incubations were terminated by freezing in -80°C ethanol, and samples were stored frozen until further treatment. Sediment cores for vertical profiles of sulfate reduction were injected with 4 µl carrier-free \( {^{35}}S\)-SO\(_4^{2-}\) (200 kBq) through each of 10 silicone-filled injection ports (1 cm apart) along the core liner. Sediment cores were incubated in darkness for 5 h at 18°C, and incubations were terminated as described above.

The radial burrow samples were extruded stepwise while still frozen, and cut in millimeter sections: 1 mm intervals to 4 mm depth, 2 mm intervals to 10 mm depth, followed by 5 mm intervals to 25 mm depth (the direction from the interface of the burrow wall and horizontally into the ambient sediment is here designated as depth). Each sediment portion was immediately transferred to a pre-weighed 15 ml centrifuge tube containing 5 ml of 5% ZnAc; the tube and sediment was then re-weighed. The vertical 2.9 cm i.d. sediment cores were sectioned into 0.5 cm segments for the upper 2 cm, and into 1 cm segments for the 2 to 10 cm depth section. About 1 g of sediment from each section was fixed with 8 ml of 5% ZnAc in centrifuge tubes. The tubes were shaken vigorously, and centrifuged at 3000 rpm (1200 \( \times \) g) for 5 min. Of the supernatant, 100 µl was assayed for \( {^{35}}S\)-SO\(_4^{2-}\), and 1.5 to 4 ml were used for turbidometrically determination of total sulfate concentration (Tabatabai 1974). The sediment was transferred quantitatively into a 100 ml reaction flask with 8 ml of 50% ethanol. Radiolabeled sulfide and total reduced inorganic sulfur (TRS) were recovered by a single-step distillation procedure (King et al. 1985, Fossing & Jørgensen 1989). After sparging for 10 to 15 min with a stream of nitrogen, 15 ml of 1 M CrCl\(_3\) in 1 M HCl was added. Sulfide liberated by a 60 min boiling distillation was trapped as ZnS in 6 ml of 5% ZnAc. Aliquots (50 to 500 µl) of the traps were subsampled for sulfide analysis (TRS) using the methylene blue technique of Cline (1969). Radioactivity in the supernatant (\( {^{35}}S\)-SO\(_4^{2-}\)) and the traps (\( {^{35}}S\)-CH\(_2\)S) were determined in a LKB 1218 scintillation counter (LKB, Pharmacia, Inc., Gaithersburg, MD, USA).

**Solid phase iron.** Sediment samples for radial profiles of solid phase iron were obtained from *Mya arenaria* burrows in the same intervals as described for the sulfate reduction assay. Solid phase iron was determined by HCl- and hydroxylamine-extractions using a method modified from Lovley & Phillips (1987). Briefly, a 100 mg subsample of homogenized sediment from each section was quickly transferred into 5 ml of 0.25 M HCl ([HCl-extractable Fe(II)]). A similar subsample was transferred into 5 ml of 0.25 M hydroxylamine hydrochloride in 0.25 M HCl ([HCl-extractable total Fe]). Both mixtures were shaken at regular intervals for 5 h at 18°C, and incubations were terminated as described above.

**Sulfate-reducing bacteria (MPN assay).** Sulfate-reducing bacteria were enumerated by the most-probable-number (MPN) technique (American Public Health Association 1969). Anoxic defined mineral medium was prepared as described by Widdel & Bak (1991), using a trace element solution described by Widdel (1986). A filter-sterilized vitamin solution (Widdel & Pfennig 1981) was added after autoclaving the medium. The pH was adjusted to 7.2, and the medium was dispensed into 100 ml airtight screw-cap bottles. The electron donors formate and acetate were added from concentrated sterile stock solutions before inoculation (final concentration 3 mM acetate and 12 mM formate). The medium was dispensed in 9 ml portions into 15 ml tubes and gassed immediately with N\(_2\)-CO\(_2\) (87:13 vol/vol). Tubes were sealed with butyl stoppers. Substrate-free medium were inoculated with wall sediment (1–3 mm) and ambient sediment (25–30 mm) from *Mya arenaria* burrows (1:10 dilution). Sediment from 8 individual burrows was pooled. The inoculum was vigorously homogenized, and 1 ml was transferred to the first tube in each dilution series (Bak & Pfennig...
and gently mixed and gassed with N₂-N₂\textsubscript{2}O. The dilution procedure was repeated serially to yield a final dilution of 10⁻⁸. An electron donor-free dilution series served as control. Three parallel dilution series were made for each treatment type. The tubes were incubated at 30°C in the dark. After 12 wk tubes were analyzed for sulfide production with CuSO₄-HCl reagent (Cord-Ruwisch 1985), and the MPN of sulfate-reducing bacteria estimated from statistical tables (American Public Health Association 1969).

**Microbial biomass.** The content of phospholipids was measured as phospholipid phosphorus according to Findlay et al. (1989) and Dobbs & Findlay (1993), and used as an estimate of total microbial biomass in burrow wall (1-3 mm) and ambient (25-30 mm) sediment. Sediment samples from several *Mya arenaria* burrows were obtained as previously described and pooled to obtain an appropriate mass. Briefly, lipids were recovered from 1 cm² sediment samples by extraction in a dichloromethane-methanol mixture. Phospholipid phosphate was liberated by potassium persulfate digestion of the extract, and measured colorimetrically (Findlay et al. 1989). The concentration of phosphorus was calculated using a standard curve prepared by digesting parallel samples with known concentrations of glycerol phosphate.

**Sediment characteristics.** Additional samples from *Mya arenaria* burrows were taken to determine sediment characteristics. Density was obtained from the wet weight of a known sediment volume. Water content was determined by drying sediment samples at 110°C for 12 h. Organic content was measured on dried sediment samples as loss-on-ignition (LOI) at 520°C for 6 h, and as particulate organic carbon (POC) and nitrogen (PON) using a Carlo Erba EA1108 Elemental Analyzer according to the procedure of Kristensen & Andersen (1987).

**Artificial burrow experiment (ABEX).** The effect of burrow irrigation on sulfate reduction and pools of reduced sulfur was assessed in an experimental study using artificial burrows in laboratory microcosms. The ABEX design provided a well-suited tool for examining the impact of burrow irrigation, because all other influences associated with burrow-dwelling animals, such as mucus secretions, were eliminated. For simplicity, the artificial burrow was not designed to mimic a *Mya arenaria* burrow, but rather to model a simple cylinder-shaped 'standard burrow'.

Surface sediment (0 to 8 cm depth) was sieved through a 1.0 mm mesh to remove macrofauna and larger particles (e.g. shell debris, gravel etc.). The sediment was allowed to settle for 3 d and was homogenized before it was transferred into acrylic core liners (20 cm long, 7.6 cm i.d.) to a depth of about 13 cm. Each core was equipped with four 14 mm i.d. sampling ports placed pairwise at 4 and 8 cm depth below the sediment surface (Fig. 2). Sampling ports were plugged tightly with rubber stoppers before the sediment was added. The sediment cores were placed in a darkened seawater tank (180 l) supplied with running oxic seawater (temperature 15 to 17°C). The sediment was allowed to equilibrate for 3 wk before introduction of the artificial burrows.

Artificial burrows (Fig. 2) were constructed using polyethylene filtration screens (Spectra/Mesh; mesh size 114 µm). A rectangular (12 X 3 cm) sheet was fused into a cylinder-shaped burrow (i.d. 8 mm) using a 'Tefal Jet Line 30' iron and sealed at the bottom with the cut-off tip of a microcentrifuge tube. The burrow was mounted on a glass rod and carefully inserted vertically.
into the center of the sediment core; the glass rod was subsequently pulled out. Seawater was passed through a Pasteur pipette to the bottom of the burrow at a constant flow rate of 100 ml h⁻¹, using a peristaltic pump. This was considerably lower than previously reported pumping rates of *Mya arenaria* (0.7 to 3.0 l h⁻¹ g⁻¹ dry tissue wt; Jorgensen & Riisgård 1988), but within the range of ventilation rates recorded for a number of other infaunal burrow dwellers (e.g. Foster-Smith 1978, Kristensen 1988). One artificial burrow was established in each of 4 sediment cores, and another 4 sediment cores with no burrows served as controls.

Sulfate reduction rates were measured after operating the cores for 6 wk after introduction of artificial burrows. The sediment cores were removed from the tank and the overlying water siphoned off. A glass rod was carefully mounted in the burrow before sampling to stabilize the burrow and diminish disturbance. Sediment subsamples were obtained by removing the sample port plugs and carefully inserting a 5 cm³ 'cut-off' syringe horizontally into each sample port towards the core center. Sampling from 2 parallel ports was done synchronously; care was taken to insert the syringe all the way to the artificial burrow to obtain the sediment adjacent to the artificial burrow wall. Sampling was conducted from top to bottom, starting with the 4 cm depth samples. Sampling of control cores (no burrows) was accomplished similarly, but without insertion of a glass rod. The syringe cores were immediately sealed with butyl septa, and each was injected with 10 µl of carrier-free ³⁵S-SO₄²⁻ (~600 kBq) within 10 min after sampling. Incubation, sediment sectioning and analysis of sulfate reduction were conducted as described previously for the *Mya arenaria* burrow samples. Samples from control cores, however, were sectioned in 5 mm intervals.

### RESULTS

**Sediment cores**

**General observations**

Sediment cores collected for measurements of vertical profiles of sulfate reduction had a brownish oxidized zone in the upper 2 to 3 cm. Below this depth, the reduced sediment was greyish-black. Visually the sediment appeared heterogeneous with an irregular transition between oxidized and reduced sediment. Faunal activity (other than *Mya arenaria*) in the sediment was evident as local extensions of the oxidized zone into the otherwise reduced sediment. Benthic microalgae were abundant on the sediment surface, whereas macroalgae (e.g. *Enteromorpha* sp., *Chaetomorpha* sp.) were more scattered in occurrence.

**Sulfate reduction and reduced inorganic sulfur**

Considerable variability was observed in vertical profiles of sulfate reduction rates (SRR) and TRS among 3 replicate sediment cores, as evidenced by relatively large standard deviations particularly near the sediment surface (Fig. 3). Rates of sulfate reduction (Fig. 3A) increased rapidly with depth from about 125 to 350 nmol cm⁻³ d⁻¹ in the upper 0–1.5 cm interval. Maximum average rates close to 350 nmol cm⁻³ d⁻¹ occurred in the 1.5–3 cm depth interval. Below 3 cm, SRR gradually declined with depth, reaching a level in the 9 to 10 cm interval (110 nmol cm⁻³ d⁻¹) similar to the rate observed in the uppermost sediment layer (0–0.5 cm).

TRS concentrations (Fig. 3B) increased gradually with depth from about 43 µmol S cm⁻³ in the upper

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**Fig. 3.** Vertical distribution of (A) sulfate reduction rates (SRR) and (B) total reduced inorganic sulfur (TRS). Values presented as mean (±SE; n = 3)
0.5 cm to about 85 to 90 \mu mol S cm\(^{-2}\) in the 2–3 cm interval. Concentrations between about 2 to 7 cm depth were more or less constant, but were followed by an increase, reaching maximum values (105 to 110 \mu mol S cm\(^{-2}\)) in the lower 7 to 10 cm.

**Mya arenaria** burrows

**General observations**

Visual observations during sampling showed that *Mya arenaria* burrows were surrounded by a 2 to 5 mm thick oxidized zone. The thickness of this zone varied considerably between individual burrows, but decreased only slightly with depth in each individual burrow. The clams were typically positioned at 12 to 20 cm depth, and only a thin oxidized sediment zone (about 0.5 to 1 mm) was evident around the shells of the animals.

**Sediment characteristics**

The presence of *Mya arenaria* burrows affected the sediment water content, which was about 24% higher in sediment immediately surrounding the burrow compared to the ambient sediment (20–25 mm depth) (data not shown). The organic content (Fig. 4A & B) was relatively high in the inner 1–3 mm region surrounding the burrow (1.05 to 1.13 % POC and 0.10 to 0.11 % PON), and decreased gradually by 10 to 15% with distance from the burrow wall. The corresponding molar C:N ratio (Fig. 4C) was relatively low in the inner 1–3 mm region (11.6 to 11.9) and increased considerably in the succeeding 2 intervals (12.9 to 13.2). In ambient sediment (>6 mm), the C:N ratio ranged from 12.2 to 12.8.

**Sulfate reduction and reduced inorganic sulfur**

Radial profiles of SRR in sediment around *Mya arenaria* burrows varied both with distance from the wall and between individual burrows (Fig. 5). In general, 2 basic types of profiles were observed. In one type (Type I, Fig. 5A–E), the highest rates were found in the inner 0–5 mm radial region followed by a gradual decline with distance, reaching a relatively constant level in the ambient (>10 mm) sediment. Maximum rates measured in the burrow wall of Type I were typically 2 to 2.5 times higher than rates in the ambient sediment, but the absolute magnitude of SRR differed markedly from burrow to burrow (e.g Fig. 5A–D). In the other type (Type II, Fig. 5F–I), maximum rates were observed in the 4–10 mm interval, the maxima were about 2-fold higher than rates in the 1–4 and 10–30 mm intervals. Of the total burrow samples, about half were characterized as Type I. There was no relation between types of profiles and the time of sampling (June and November).

The radial profiles of TRS varied considerably among samples and within profiles (Fig. 6). TRS concentrations generally increased with distance from the burrow wall, although some profiles were erratic and irregular (e.g. Fig. 6B, C & I) with more or less constant levels among depth intervals. There was no general relationship between SRR and the corresponding TRS concent-
Fig. 5. Radial burrow profiles of sulfate reduction rates (SRR) in sediment surrounding Mya arenaria burrows. Each profile represents 1 individual burrow sample obtained from 5 to 6 cm depth. Values are single determinations of each radial depth. [D–F] represent burrows sampled in November, the remaining burrows were sampled in June. Note different scaling.

Fig. 6. Radial burrow profiles of total reduced inorganic sulfur (TRS) in sediment surrounding Mya arenaria burrows. Each profile represents 1 individual burrow sample obtained from 5 to 6 cm depth. Profiles are equivalent to those in Fig. 5 except for [D–F] (o) which represent samples without corresponding SRR data. Note different scaling.
SULFATE REDUCTION

![Graph A: Sulfate Reduction Rates](image)

![Graph B: Reduced Inorganic Sulfur](image)

![Graph C: Turnover Time](image)

**Fig. 7.** Radial profiles of (A) sulfate reduction rates (SRR), (B) total reduced inorganic sulfur (TRS) and (C) turnover time of reduced sulfides, in sediment surrounding *Mya arenaria* burrows. Values are mean (±SE; n = 9) of profiles presented in Figs. 5 & 6.

Table 1. Cell numbers of sulfate-reducing bacteria (SRB) and phospholipid phosphorus concentrations in sediment surrounding *Mya arenaria* burrows. Enumeration of SRB was made by the most-probable-number (MPN) technique. Phospholipid-phosphorus content was used as an estimate of total microbial biomass. See 'Microbial numbers and biomass' for details.

<table>
<thead>
<tr>
<th>Burrow wall sediment</th>
<th>SRB (cells cm⁻³ sediment)</th>
<th>Phospholipid phosphorus (nmol P g⁻¹ dry wt, ± SE)</th>
</tr>
</thead>
<tbody>
<tr>
<td>(0–3 mm)</td>
<td>1.1 × 10⁸</td>
<td>149.2 ± 18.5</td>
</tr>
<tr>
<td>(25–30 mm)</td>
<td>1.1 × 10⁵</td>
<td>87.3 ± 11.0</td>
</tr>
</tbody>
</table>

Concentrations by the corresponding sulfate reduction rates provided a simple estimate of the sulfur dynamics (King et al. 1985). $T_r$ increased gradually with distance from the burrow wall (Fig. 7C) to values about 2.2 times more rapid in the burrow wall (0–4 mm) than in ambient (15–25 mm) sediment.

Solid phase iron

The pool of reduced HCl-extractable particulate iron, Fe(II)_{HCl}, increased gradually with distance from the burrow wall, and was about 1.3 times higher in ambient (10–25 mm) sediment relative to burrow wall (0–3 mm) sediment (Fig. 8). Fe(III)_{HCl}, determined as the difference between the total HCl-extractable Fe content and Fe(II)_{HCl} (illustrated as the area between the 2 profiles in Fig. 8), was highest in the innermost zone around burrows (29.0 μmol cm⁻³) and declined rapidly with distance. The total extracted Fe content was only markedly different from the Fe(II)_{HCl}-extractable iron concentrations.
tracted Fe in the innermost 3 mm, indicating that Fe(III)$_{red}$ was present adjacent to the burrow wall only.

Microbial numbers and biomass

MPN determinations of acetate/formate-utilizing sulfate-reducing bacteria were about a factor of 10 higher in the burrow wall (0–3 mm) sediment as compared to the ambient (25–30 mm) sediment (Table 1). Total microbial biomass, estimated as the content of phospholipid phosphorus, was 1.7 times greater in the burrow wall (0–3 mm) than in ambient (25–30 mm) sediment (Table 1).

Artificial burrow experiment (ABEX)

General observations

The sediment cores developed a 6 to 8 mm thick oxidized brownish zone in the topmost sediment within the first 2 wk of incubation. Below this, the sediment had a greyish-brownish colour down to about 2 cm depth followed by a reduced, greyish-black appearance in the lower sediment column. There was an irregular transition between oxidized and reduced sediment, with no visual difference between sediment cores with artificial burrows and control cores. In the case of the artificially irrigated burrows, however, there was an extension of the oxidized zone in the center as evidenced by a 1 to 3 mm thick brownish oxidized layer surrounding the burrows. During sampling for SRR assays, some of the oxidized sediment immediately adjacent to the artificial burrow wall may have been lost. The originally white polyethylene burrow material developed an orange-reddish colour in the course of the experiment, most likely from the precipitation of oxidized iron compounds.

Sulfate reduction and reduced inorganic sulfur

Irrigation of the artificial burrows had a pronounced impact on both SRR and TRS (Fig. 9). Results from the 4 and 8 cm depth samples showed the same pattern, and have been combined in the figures. Artificial burrow irrigation suppressed sulfate reduction markedly in sediment surrounding burrows (77.8 to 91.1 nmol cm$^{-3}$ d$^{-1}$ in the 0–6 mm region), and SRR generally increased with the distance from the burrow wall, in the 15–25 mm region approaching values similar to control cores (Fig. 9A). SRR in control cores averaged 169.3 ± 6.1 (SE) nmol cm$^{-3}$ d$^{-1}$ and was on average a factor of 2 higher than in artificially irrigated cores in the 0–6 mm region. TRS concentrations (Fig. 9B) in artificially irrigated sediment cores increased gradually with distance from the burrow wall, reaching values about 1.5 to 1.6 times higher in the 20–25 mm interval relative to the 0–1 mm interval. Concentrations in control cores averaged $87.6 \pm 1.5$ (SE) $\mu$mol S cm$^{-3}$. TRS concentrations in irrigated cores approached concentrations in the control cores in the 20–25 mm interval.

DISCUSSION

The rates of sulfate reduction measured in this study lie within ranges previously reported for subtidal and intertidal coastal sediments (e.g. Skyring 1987, Mackin...
bivalves into the surrounding sediment could enhance burrow wall may turn anoxic during the incubation
nate, acetate and propionate, excreted by these
1994). Contrary to this, part of the oxic zone near the
1990). Short-chain organic acids, such as lactate, succi-
injection of oxygen into the burrow sediment
method may seriously underestimate the true rate of
S1 Accumulation of Fe(II) has previously been shown in sediment sur-
S1 Sulfate reduction in oxidized sediments has been
demonstrated in the presence of oxygen (Canfield & Des
10: 181-194, 1996
Aquat Microb Ecol
local growth of SRB and sulfate reduction in burrow sediments (Bussmann & Reichardt 1991). However, further studies are needed to support this hypothesis.

The 2 basic types of sulfate reduction profiles. Type I with maximum rates in the 1 to 5 mm radial layer and, Type II with highest rates in the 4 to 10 mm radial region may be attributed to inter-burrow differences in age, sediment structure and composition, and to variations in bivalve ventilation and metabolic activity. *Mya arenaria*, in the study, site, is harvested commercially and new burrows are constantly being established. Relatively old burrows may be richer in both organic content and microbial populations than newly established burrows. Type II burrows with lowest TRS concentrations in the burrow wall sediment appear more oxidized than Type I burrows. SRR may, therefore, be suppressed in the first mm intervals of Type II burrows, displacing the SRR peak radially away from the burrow wall. It is not clear whether this radial SRR displacement in Type II burrows is due to suppression of SRR in the innermost burrow wall sediment, or whether this is an artifact resulting from sulfide reoxidation in this zone during incubation, or is due to other factors. As noted earlier, the radiotracer assay for SRR may underestimate rates in oxidized burrow wall sediment due to sulfide reoxidation during incubation. Differences in bivalve ventilation activity are likely to influence the extent of oxidation of the burrow sediment. Previous studies (Pedersen 1992) have reported high temporal variation in metabolic activity of *M. arenaria* without any direct relation to the tidal pattern. Thus, the Type II profile may represent burrows inhabited by relatively active clams. In contrast to the pattern for *M. arenaria* burrows, SRR was lowest in sediments adjacent to the artificial burrows. This suggests that the supply of oxygen via irrigation was sufficient to cause a shift in the relative importance of organic matter degradation by SRR. The results further emphasize the potential role of macrofauna in enhancing organic matter availability adjacent to burrows and suggest that substrate enrichment is associated with the activities of *M. arenaria*.

In spite of the variability among samples, TRS concentration profiles indicate that sulfur is lost from the sediment, possibly by oxidation of reduced sulfur species. The increasing turnover time of reduced inorganic sulfur compounds with distance from the burrow wall (Fig. 7C) suggests that *Mya arenaria* burrows play an important role in the oxidation of sulfides and sulfide removal from the sediment. The profiles of TRS obtained in the artificial burrow experiment similarly indicate loss of sulfides from the sediment surrounding the irrigated burrows. Sulfide oxidation was most likely mediated by the artificial irrigation of oxygen-rich water to the otherwise reduced sediment. Reduced sulfides may be oxidized by various chemical and bacterial processes (Jorgensen 1983, 1988). Besides oxidation by O2 in the oxic sediment, oxidation processes in the anoxic sediment zones may be mediated by nitrate, manganese or iron as electron acceptors (Aller & Rude 1988, Fossing & Jorgensen 1990, King 1990, Elsgaard & Jorgensen 1992).

Phospholipid phosphorus concentrations indicated that total microbial biomass was elevated in burrow walls of *Mya arenaria* relative to ambient sediment. The same pattern has been shown for tubes and burrows of other macrofaunal organisms (e.g. Aller & Yingst 1978, 1985, Dobbs & Guckert 1988). The higher microbial biomass in the innermost burrow wall may also be reflected in the relative lower C:N ratios observed in this zone, which suggest a nitrogen-rich microbial contribution to total sedimentary organic matter. The phospholipid phosphorus concentrations reported here are within the range found by Findlay et al. (1989) for Lowes Cove sediment and by others for a variety of estuarine sediments (e.g. White et al. 1979, Federle et al. 1983). Using a conversion factor $3.43 \times 10^6$ cells per 100 nmol of phospholipid phosphorus (Findlay et al. 1989), the estimated cell densities in the present study are $5.0 \times 10^6$ and $3.3 \times 10^6$ cells cm$^{-3}$ of sediment in burrow wall and ambient sediment, respectively. Cell numbers in this range are common for intertidal sediments (e.g. Dale 1974, Rublee 1982, DeFlaun & Mayer 1983).

Using a biochemical analysis of phospholipid ester-linked fatty acids (PLFA), Dobbs & Guckert (1988) demonstrated higher densities of SRB, such as *Desulfovibrio* spp., in the relatively aerobic burrow lining of *Callianassa trilobata*, a deep-burrowing thalassinid crustacean, compared to ambient sediment. In the present study MPN results also demonstrated a higher abundance of SRB in burrow wall sediment than in ambient sediment. This pattern corresponded with the distribution of sulfate reduction rates. The numbers of SRB obtained here ($1.1 \times 10^7$ cm$^{-3}$ and $1.1 \times 10^8$ cm$^{-3}$ in ambient and burrow wall sediment, respectively) are higher than values previously reported on the basis of viable counting techniques ($10^3$ to $10^6$ cm$^{-3}$ of sediment, e.g. Jorgensen 1978b, 1983, Hines & Jones 1985, Gibson et al. 1987, Jorgensen & Bak 1991, Parkes et al. 1993). The estimated specific metabolic rate of SRB in burrow wall (0 to 4 mm) and ambient (20 to 25 mm) sediment was $1.72 \times 10^{-8}$ and $9.27 \times 10^{-8}$ mmol SO$_4^{2-}$ cell$^{-1}$ d$^{-1}$, respectively, which is within the range reported for SRB growing in pure cultures ($10^{-8}$ to $10^{-5}$ mmol SO$_4^{2-}$ cell$^{-1}$ d$^{-1}$; Jorgensen 1978b). This suggests that the MPN results in this study provided realistic estimates of true cell numbers.

The present study illustrates the effects of *Mya arenaria* and its burrows on biogeochemical activities...
related to the sulfur cycle in an intertidal sediment. The results suggest that a zone of elevated sulfur transformations occurs in close proximity to burrow walls. Sulfate reduction rates were up to 2.5 times higher in the inner 1 to 5 mm region surrounding the burrows, whereas pools of reduced sulfur increased with the distance from burrow walls to values ~1.5 times higher than in ambient sediment. The radial profiles of SRR indicated that M. arenaria burrows typically affect SRR in surrounding sediment within a distance of roughly 8 mm from the burrow wall, corresponding to 4.5 cm² of sediment, assuming an average burrow diameter of 10 mm. For typical densities of 200 burrows m⁻², about 9% of the total surface area is affected by M. arenaria. Failure to account for enhanced activity in M. arenaria burrows could result in an underestimate of SRR by about 15% relative to bulk sediment rates. If patterns for other macrofauna are similar to those of M. arenaria, sulfate reduction near burrow walls may account for a substantial fraction of the total activity. In the ABEX experiment, artificially irrigated burrows suppressed SRR in the innermost zone around burrows, whereas radial profiles of TRS generally resembled those from M. arenaria burrows in situ, indicating loss of reduced sulfur compounds from the burrow wall. Substrate enrichment in the burrow wall sediment is most likely responsible for the enhanced SRR around M. arenaria burrows. More detailed studies are required to clarify and identify the origin and the mechanisms involved in such substrate enrichment. In addition, the relationship between M. arenaria behaviour, e.g. ventilation patterns and metabolic activity, and the patterns and the dynamics of SRR needs greater attention.

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